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PREPARATION OF EMISSION INVENTORIES OF TOXIC AIR CONTAMINANTS FOR THE BAY AREA

FINAL REPORT 2 STI-906020.07-2771-FR2

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1. INTRODUCTION

The Bay Area Air Quality Management District (BAAQMD) is carrying out the Community Air Risk Evaluation (CARE) program to characterize and reduce health risks from toxic air contaminants (TACs) emitted in the San Francisco Bay Area (SFBA). In support of the CARE program's goals, screening-level gridded emission inventories of TACs were prepared (including diesel exhaust). These screening-level inventories were assembled top-down from readily available information and represent an early step from which subsequent CARE program activities may be carried forward. They are intended to aid in the identification and prioritization of further inventory development activities, to support initial exposure modeling runs, and to facilitate selection of a study community in the SFBA. Although the screening-level inventories represent many of the TACs and emissions sources that are important in the SFBA, they should be considered "working versions under development"—i.e., inventories that are useful, but should be augmented and improved in consideration of reviewers' feedback or as additional inventory data are developed bottom-up through CARE program activities.

Mass-based and risk-weighted annual emission inventories of TACs were prepared by acquiring and processing existing and available information sources, including year-2000 emissions data previously prepared by BAAQMD. BAAQMD staff provided Sonoma Technology, Inc. (STI) with a county-level area and non-road mobile source inventory of emissions of total organic gases (TOG) and particulate matter less than 10 microns (PM_{10}). These emissions were back-cast to year 2000 from a base year of 2002. In addition, BAAQMD staff generated year-2000 on-road motor vehicle emissions using the California Air Resources Board's (ARB) EMFAC2002 model (version 2.2). These emissions were allocated to the BAAQMD's 2-km \times 2-km modeling grid using the California Department of Transportation's (Caltrans) Direct Travel Impact Model Version 4 (DTIM4) and year-2000 SFBA travel network data provided by the Metropolitan Transportation Commission (MTC). BAAQMD produced summer and winter average weekday inventories for on-road mobile sources, and these two inventories were averaged by grid cell to produce an annualized on-road emissions inventory. Finally, BAAQMD provided STI with a year-2000 inventory of TAC emissions from point sources in the BAAQMD, data that were compiled from emissions reported by individual facilities.¹ (STI also obtained a year-2000 inventory of point source TOG and PM_{10} emissions in the Bay Area from the ARB for purposes of comparison.)

Other information sources used to prepare TAC emission inventories included (1) chemical speciation profiles published by the U.S. Environmental Protection Agency (EPA), ARB, and Desert Research Institute (DRI); (2) geographic information systems (GIS) databases acquired from the Association of Bay Area Governments (ABAG) and other sources; (3) ratios of particulate matter less than 2.5 microns $(PM_{2.5})$: PM_{10} calculated from information published

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¹ Facility emission reports were provided during calendar year 2000 and, in most cases, represent the most recent 12-month period. Therefore, because the reporting dates for various facilities are staggered throughout the year, some reported emissions include those from the previous calendar year. However, these reported emissions were judged to be sufficiently representative of year-2000 for purposes of this project.

by the ARB; and (4) inhalation unit risk factors and reference concentrations available from the ARB and EPA. The resultant inventories should be considered suitable for initial exposure modeling and data analysis with a view toward prioritizing TACs and emissions source categories of concern for future research efforts. Conclusions and recommendations for further improvements to the inventories are presented in Section 6.

2. SUMMARY OF EXISTING INVENTORIES

The BAAQMD's pre-existing inventories of TOG, PM_{10} , and TACs were used as the basis for compiling annualized, screening-level TAC emission inventories for the CARE program. Figure 2-1 illustrates the distribution of total TOG and PM_{10} emissions by major source category (point, on-road mobile, and area/non-road). Section 2 discusses each major source category in greater detail and outlines the techniques used to geographically allocate emissions to the BAAQMD's 2-km \times 2-km modeling grid (see Appendix A).

Figure 2-1. 2000 TOG and PM_{10} emissions by major source type for the BAAQMD.

2.1 EMISSIONS FROM AREA AND NON-ROAD MOBILE SOURCES

The BAAQMD provided STI with county-level area and non-road mobile source inventories of year-2000 TOG and PM_{10} emissions for 356 emission inventory codes (EICs). These inventories incorporated the latest emission estimates prepared by the BAAQMD, including the recent improvements to the area and non-road mobile source inventories listed below:

- Ship emission estimates were updated to include emissions occurring within 100 miles of county shorelines (previous estimates only included emissions occurring within 3 miles of county shorelines).
- Emission estimates for prepared for cargo handling equipment at the Port of Oakland.
- Improvements were made to the distribution of residential wood burning emissions among SFBA counties based on the results of a wood burning survey (True North Research, 2005). Figure 2-2 shows an emission density plot of current PM_{10} emissions from residential wood burning.
- County-level estimates of emissions from livestock waste were updated.

Figure 2-2. Emission density plot of PM_{10} emissions from residential wood burning.

Emission estimates for $PM_{2.5}$ were developed by applying $PM_{2.5}$ -to- PM_{10} size fractions recommended by the ARB for each EIC code (see Appendix B). Figures 2-3 and 2-4 summarize area and non-road mobile source TOG, PM_{10} , and $PM_{2.5}$ emissions by detailed source categories, and Figure 2-5 shows an emission density plot of area and non-road mobile source TOG emissions (note that BAAQMD boundaries include only the western portion of Solano County and the southern portion of Sonoma County). Important sources of TOG include evaporative sources (e.g., petroleum marketing and consumer products) and combustion sources (e.g., off-road equipment and residential fuel combustion). Sources of TOG associated with the decomposition of organic matter (e.g., livestock waste and landfills) emit most TOG in the form of methane—a non-toxic gas. Important sources of PM include combustion sources (e.g., residential fuel combustion and off-road equipment). Sources of fugitive dust emit most PM in the form of non-toxic geologic material.

Figure 2-3. Area and non-road mobile source TOG emissions (tons/day) by source category.

Figure 2-4. Area and non-road mobile source PM emissions (tons/day) by source category.

Figure 2-5. Emission density plot of area and non-road mobile source TOG emissions.

County-level area and non-road mobile emissions were geographically distributed using GIS databases. GIS databases with suitable spatial resolutions were selected as surrogates for the locations of emissions sources. Countywide emissions were allocated to individual grid cells proportionally according to the spatial patterns of the surrogate GIS data. Examples of some GIS databases employed for the process include land use area (e.g., industrial land use), line length

(e.g., railroad track length), line density (e.g., roadway traffic activity), and point count (e.g., number of dry cleaning locations). Allocation factors were developed for individual grid cells by processing GIS data within a customized ArcGIS Visual Basic (VBA) program that outputs allocation factors by grid cell to Microsoft Access database tables. Four basic types of spatial allocation calculations were used to develop the spatial surrogates applied to the inventory, based on the type of GIS data (i.e., polygon, line, or point).

A variety of GIS data sets were used to geographically distribute countywide emissions. The most often-used data set for this process was the Existing Land Use – 2000, developed by the ABAG. The ABAG land use database incorporates the U.S. Geological Survey National Land Cover Dataset (NLCD) as well as county assessors' data on land use. Another important data source was the Central California Ozone Study (CCOS) gridded surrogates project (Funk et al., 2001) that included representations of the geographic locations of businesses derived from records of business addresses. Addresses for auto body shops, dry cleaners, and variety of other types of businesses were geocoded to estimate geographic locations, which Funk et al. (2001) used in turn to calculate gridded spatial allocation factors. In total, 46 surrogates were developed for spatially allocating the area and non-road sources. The details about GIS data sources and methods used for spatial allocation are provided in Appendix C.

2.2 EMISSIONS FROM ON-ROAD MOBILE SOURCES

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The BAAQMD provided STI with gridded on-road mobile source inventories of TOG and PM10 emissions in Modeling Emissions Data System (MEDS) format. BAAQMD staff generated year-2000 on-road mobile source emission estimates using the ARB's EMFAC2002 model (version 2.2). These emissions were allocated to $BAAQMD's 2-km \times 2-km$ modeling grid using Caltrans' DTIM4 and year-2000 SFBA travel network data provided by the MTC. The final emission inventories provided to STI incorporated a number of recent improvements that BAAQMD staff made to on-road mobile source emission estimates:

- The redistribution of heavy-duty truck emissions from a portion of Route 580 to I-880.²
- The incorporation of new emission estimates for trucks operating within the Port of Oakland.
- A vehicle speed correction for a stretch of Highway 152 east of Gilroy that eliminated a previous overestimate of emissions for that region.
- Diesel PM emissions from heavy-duty vehicles were updated based on results from a new draft version of ARB's EMFAC model which uses truck survey results and global positioning system (GPS) techniques to estimate vehicle miles traveled (VMT). This methodology is more realistic than the one employed in the current version of EMFAC (Version 2.2), which relies primarily on registration data to estimate VMT.

BAAQMD generated summer and winter MEDS files for 2000, and these two inventories were averaged by grid cell to produce an annualized on-road emissions inventory.

² Heavy-duty trucks over 9,000 pounds are prohibited on Route 580 from Foothill Blvd. in San Leandro to Grand Avenue in Oakland. As a result, most of these trucks use I-880 instead, a traffic feature that was not captured in previous gridded inventories.

As with emissions from area and off-road mobile sources, emission estimates for $PM_{2.5}$ were developed from the on-road PM_{10} inventory by applying ARB $PM_{2.5}$ -to- PM_{10} size fractions recommended by the ARB. Figures 2-6 and 2-7 summarize on-road mobile source TOG, PM_{10} , and PM2.5 emissions by source category, and Figure 2-8 shows an emission density plot of the gridded on-road mobile source TOG inventory provided by BAAQMD. Evaporative losses are important sources of TOG and exhaust emissions are important sources of both TOG and PM.

Figure 2-6. On-road mobile source TOG emissions (tons/day) by source category.

Figure 2-7. On-road mobile source PM emissions (tons/day) by source category.

Figure 2-8. Emission density plot of on-road mobile source TOG emissions.

2.3 EMISSIONS FROM POINT SOURCES

BAAQMD provided STI with a year-2000 inventory of TAC emissions from point sources within the district, data that were compiled from emissions reported by individual facilities. However, for purposes of direct comparison to the criteria-pollutant data summarized in Sections 2.1 and 2.2, STI downloaded year-2000 TOG, reactive organic gases (ROG) and PM₁₀ point-source emissions from ARB's Facility Search Engine. Figures 2-9 and 2-10 summarize point source ROG and PM_{10} emissions by facility. (ROG emissions are displayed because the TOG inventory is dominated by landfills, which emit most TOG in the form of methane—a non-toxic gas). Petroleum refineries are important sources of ROG and PM_{10} , and electrical generation facilities and landfills are also important sources of PM_{10} .

Figure 2-9. Point source ROG emissions (tons/year) by facility.

Figure 2-10. Point source PM_{10} emissions (tons/year) by facility.

In order to assign a geographical location to individual point sources, facility addresses reported in the BAAQMD's point source inventory were geocoded. The geocoding was achieved using the TeleAtlas EZ-Locate software. The TeleAtlas' geocoding process relies on a combination of postal standardization software, positionally accurate maps, and advanced geocoding techniques to generate the most accurate geocode coordinates available. Each point source in the BAAQMD's inventory file was assigned a geographic location and a geocode match type field. The match type is determined according to the accuracy of the geocoded address. A match type of "1" is the best match type possible—the geographic location is accurate to an exact house number within a single side of a single street block. A match type of "4" is accurate only to the 5-digit ZIP Code centroid. Of 3,359 unique addresses in the BAAQMD's point source inventory file, 4% (137 addresses) received a match type "4" and 96% (3,222 addresses) received a match type "1".

Following STI's geographic location assignments, BAAQMD staff reviewed spatial information for 90 of the BAAQMD's most significant point sources. Sites selected for review included all Title V facilities³ and the 20 sites with the highest toxicity-weighted emissions for either cancer, chronic or acute health risks. The BAAQMD provided STI with correct UTM coordinates for 56 of these sites and correct grid cell locations for 34 of these sites. This review also included 21 sites that STI designated a match type of "4", and it should be noted that the estimated locations of remaining point sources with a match type "4" are inexact and should be corrected or taken into consideration during exposure modeling and risk assessment efforts. Figure 2-11 illustrates the geographic distribution of benzene emissions from point sources. Details about the point source inventory geocoding match type results are documented in Appendix C.

³These facilities are required to obtain a Title V operating permit in accordance with federal Clean Air Act Amendments of 1990 based on their potential to produce certain thresholds of air pollution.

Figure 2-11. Emission density plot of benzene emissions from point sources.

3. DEVELOPMENT OF TOXIC AIR CONTAMINANT INVENTORIES

Several data processing steps were followed to prepare the inventories for chemical speciation. STI assigned speciation profiles to each EIC or source classification code (SCC) in the area, non-road mobile, and on-road mobile source TOG and PM inventories. The ARB has developed a database of TOG and PM speciation profiles, as well as a cross-reference file that indicates which TOG or PM profile is assigned to each EIC code. In most cases, speciation profile-to-EIC code assignments recommended by the ARB were followed. However, where no speciation profile (or a composite profile) was recommended for a given source category by the ARB, appropriate profiles from DRI or the EPA Speciate 3.2 database were utilized. (Appendix C contains a listing of the speciation profiles used in this study and a cross-reference table matching these profiles to each EIC/SCC code in the inventories).

Several speciation profiles contained weight fractions for chromium, a compound that is highly toxic when it exists in its hexavalent state. Data on ambient concentrations of hexavalent chromium in California indicate that hexavalent chromium comprises 3% to 8% of total ambient chromium (California Air Resources Board, 1985). Therefore, in consultation with BAAQMD staff, STI assumed that 5% of the chromium fractions in the speciation profiles used in this project was made up of hexavalent chromium; the speciation profiles were adjusted accordingly. No other revisions were made to the selected speciation profiles. It should also be noted that the BAAQMD provided STI with a year-2000 inventory of TACs from point sources in the SFBA; therefore, no processing steps were required to accomplish chemical speciation of the point source inventories.

In addition, because particulate emissions from diesel-fueled engines have been defined as a toxic air contaminant by the ARB, all sources of diesel particulate matter (DPM) were identified prior to the application of speciation profiles. This identification was made by flagging all sources assigned to a diesel exhaust speciation profile and—based on guidance from BAAQMD staff—all ship emissions.⁴ PM_{10} and $PM_{2.5}$ emissions from these sources were treated as DPM and included in the final TAC inventories.

Selected speciation profiles were used to transform TOG and PM emissions into individual chemical species so that TAC emissions from area, non-road mobile, and on-road mobile sources could be estimated. These estimates were combined with the inventory of TAC emissions from point sources provided by the BAAQMD to form a complete inventory of TACs. TAC emissions by pollutant and source category can be seen in Figures 3-1 and 3-2. Note that Figures 3-1 and 3-2 show emissions on a mass basis; however, the toxicity and the geographic distribution of each TAC must be considered in order to evaluate its potential for posing human health risks. Section 4 discusses measures of toxicity and plots the geographic distributions of several potentially important TACs.

 4 Some commercial marine vessels burn residual fuel and are assigned to ARB's "Marine Vessels – Liquid Fuel" PM speciation profile. However, BAAQMD staff determined that previous ARB studies at the Ports of Los Angeles and Long Beach treated particulate emissions from residual-fueled ships as DPM. This was done for two reasons: (1) residual-fueled ships burn a fossil fuel in diesel (compression) engines; and (2) health studies for determining risk factors were based on older diesel fuels which closely resemble the "residual fuels" of today.

Figure 3-1. Average daily TAC emissions by species.

Figure 3-2. Average daily TAC emissions by detailed source category.

4. DEVELOPMENT OF TOXICITY-WEIGHTED TOXIC INVENTORIES

To develop toxicity-weighted emission inventories, STI applied available cancer unit risk (UR) factors and non-cancer reference concentrations (RfC) for the inhalation exposure pathway. UR factors estimate the expected change in the rate of observed adverse effects per unit change in dose (or air concentration). An RfC is a regulatory definition that indicates the dose at which no adverse effects are expected plus a safety margin allowing for measurement uncertainty plus another safety margin based on professional toxicologists' judgment. UR factors and RfCs were compiled from the following information sources in declining order of preference: ARB in conjunction with EPA's Office of Environmental Health Hazard Assessment (OEHHA), the EPA's Integrated Risk Information System (IRIS), and the EPA's Technology Transfer Network. Secondary sources were used to estimate factors for important TACs not available in the preferred references. When URs or RfCs were reported as a range (e.g., Benzene, CAS# 71432, IRIS), the high end of the range was used to prepare the toxicity-weighted emission inventory, and range-related uncertainties were calculated and documented. Details about the selection, application, and calculation of quantifiable uncertainties are documented in Appendix D.

Mass-based emissions for all TACs were converted to toxicity-weighted emissions for cancer, chronic, and acute risks. Toxicity-weighted emissions are reported in units of "mass equivalents per unit time". For risks of cancer due to inhalation, the mass equivalent of a specific TAC is the estimated mass of hypothetical compound "X" that poses a cancer risk equal to that of the emitted mass of the TAC of interest (where "X" is defined as having a UR factor equal to 1 $(\mu g/m^3)^{-1}$). Thus, toxicity-weighted emissions for cancer due to inhalation exposure are calculated according to Equation 4-1.

$$
Emissions \times UR_i \div UR_X = Toxicity-Weighted Emissions \tag{4-1}
$$

where:

For non-cancer risks due to inhalation (whether acute or chronic), the mass equivalent of a specific TAC is the estimated mass of hypothetical compound "Y", which would be expected to pose a risk equal to that of the emitted mass of the TAC of interest (where "Y" is defined as having an RfC equal to unity). Thus, toxicity-weighted emissions for acute or chronic effects due to inhalation exposure are calculated according to Equation 4-2.

where:

Figures 4-1 through 4-6 show toxicity-weighted emissions by pollutant and source category for cancer-related, chronic, and acute effects caused by inhalation exposure. Diesel particulate matter (DPM) comprises 81% of cancer toxicity-weighted emissions. Constructionrelated activities, industrial and commercial equipment, and on-road mobile sources contribute almost three-fourths of the cancer toxicity-weighted emissions. Acrolein and formaldehyde appear significant when considering toxicity-weighted emissions for chronic and acute effects. On-road mobile sources and aircraft are the two most important source categories for non-cancer risks, comprising 57% of all chronic toxicity-weighted emissions and almost 80% of acute toxicity-weighted emissions.

Figure 4-1. Cancer toxicity-weighted emissions by pollutant.

Figure 4-2. Cancer toxicity-weighted emissions by source category.

Figure 4-3. Chronic toxicity-weighted emissions by pollutant.

Figure 4-4. Chronic toxicity-weighted emissions by source type.

Figure 4-5. Acute toxicity-weighted emissions by pollutant.

Figure 4-6. Acute toxicity-weighted emissions by source category.

Geographic distributions of emissions for several of the TACs prominent in Figures 4-1 through 4-6 are illustrated in Figures 4-7 through 4-14, including 1,3-butadiene, DPM, acrolein, benzene, hexane, MTBE, formaldehyde, and toluene. In addition, geographic distributions of cancer toxicity-weighted, chronic toxicity-weighted, and acute toxicity-weighted emissions are shown in Figures 4-15 through 4-17.

Figure 4-7. Emission density plot of 1,3-butadiene emissions.

Figure 4-8. Emission density plot of diesel particulate emissions.

Figure 4-9. Emission density plot of acrolein emissions.

Figure 4-10. Emission density plot of benzene emissions.

Figure 4-11. Emission density plot of hexane emissions.

Figure 4-12. Emission density plot of MTBE emissions.

Figure 4-13. Emission density plot of formaldehyde emissions.

Figure 4-14. Emission density plot of toluene emissions.

Figure 4-15. Emission density plot of cancer toxicity-weighted emissions.

Figure 4-16. Emission density plot of chronic toxicity-weighted emissions.

Figure 4-17. Emission density plot of acute toxicity-weighted emissions.

5. DEMOGRAPHIC AND HEALTH STATISTICS DATA

In addition to TAC emission inventory development, demographic and health statistics data were acquired and geographically distributed in an effort to identify the locations of at-risk or sensitive populations. Several demographic and health parameters were considered:

- Populations under 18 and over 64. Youthful and elderly populations are especially vulnerable to exposures to toxic pollutants. Exposures at early ages are more likely to lead eventually to cancer or other chronic effects than are exposures at later ages, and the elderly are more vulnerable than the average population to acute morbidity or mortality from exposures to toxic pollutants.
- Income less than 185% of federal poverty level. Co-analyses of the TAC emission inventories with indicators of economically disadvantaged populations may indicate areas in which environmental justice issues are a concern.
- Age-adjusted rates of asthma-related hospitalizations in children 14 years of age and below. Note that great caution should be used when interpreting hospital admissions data. Numerous confounding factors can interfere with the analyses, either masking true causal relationships or falsely highlighting arbitrary correlations.

Gridded population data were developed for the BAAQMD using census block-group levels of data from U.S. Census 2000 summary files. The block-group population data for each variable of interest were spatially allocated to the BAAQMD's $2-km \times 2-km$ modeling grid using spatial allocation factors that were developed for individual grid cells by processing GIS-based census block-group data within a customized ArcGIS VBA program. Census block-group populations were then multiplied by the spatial allocation factors in order to acquire grid-level population values. The gridded population data are shown in Figures 5-1 through 5-3. These data show the population variations for each demographic variable.

In addition, the BAAQMD provided STI with age-adjusted asthma hospitalization rates by ZIP Code for children ages 0-14 obtained from the California Office of Statewide Health Planning and Development (OSHPD) (Community Action to Fight Asthma, 2004). STI spatially allocated these data to the BAAQMD's $2-km \times 2-km$ modeling grid using the same ArcGIS program described above. Figure 5-4 depicts the gridded, age-adjusted asthma hospitalization rates (note that not all ZIP Codes have corresponding data).

These demographic and health statistics data sets may be used to quickly identify communities of potential concern. In addition, the population data may be used as inputs for exposure-assessment modeling. However, it should again be noted that no attempt is being made to demonstrate a causative relationship between gridded TAC emissions data and the gridded asthma rates data shown in Figure 5-4.

Figure 5-1. Population density of persons under age 18 in the BAAQMD.

Figure 5-2. Population density of persons over age 64 in the BAAQMD.

Figure 5-3. Population density of persons with a personal income of less than 185% of federal poverty level in the BAAQMD.

Figure 5-4. Asthma hospitalization rates for children age 14 and under in the BAAQMD.

6. CONCLUSIONS AND RECOMMENDATIONS

The TAC inventories summarized in this document represent the first step of inventory development for the CARE program. These screening-level inventories are useful for prioritizing further inventory development activities, conducting preliminary exposure modeling runs, and selecting communities in the SFBA for further study. In addition, these inventories in combination with demographic and health data—may be used to develop preliminary, screening-level targets for emission mitigation measures using grant and incentive funds. Although refinements to these inventories should be pursued, they represent the best available information to begin the process of targeting source types and locations for potential emission mitigation measures. This section characterizes general aspects of the TAC inventories and provides recommendations for continued inventory development.

The TAC inventories prepared to date should be treated as "working versions under development". They were prepared from readily available, pre-existing information sources—an approach that was both time- and cost-efficient. A more technically rigorous approach to inventory development—though requiring more effort and resources—is a "bottom-up" approach. Bottom-up inventory development involves gathering observations of emissions or supporting data directly from emissions sources. Examples of bottom-up methods include direct measurements of emissions at their sources; surveys of emissions-related activities by telephone, mail, or in person; or measurement of emissions-producing activities with monitoring devices, such as automated daily traffic counters or meters that measure fuel consumption.

The TAC inventories developed to date have a number of strengths, but also several potentially important weaknesses. Strengths of the inventories include

- use of the most up-to-date criteria pollutant inventories for area, non-road mobile source, and on-road mobile source emissions, inventories which incorporated the refinements described in Section 2 (such as corrections to the geographic distribution of heavy-duty truck emissions);
- TAC emissions for on-road mobile sources, which were estimated using EPArecommended methods but were improved with California-specific chemical speciation data;
- TAC emissions for point sources that were directly reported, many of which are based on bottom-up approaches; and
- spatial allocations of emissions, which produced inventories that are accurate and spatially well-resolved relative to the resolution of the modeling grid $(2 \text{ km} \times 2 \text{ km})$.

However, some TACs are likely omitted from the inventory because they are not components of TOG or PM_{10} (e.g., hydrofluoric acid) or because they are infrequently measured for chemical speciation profiles. Two examples of omitted TACs—quinoline (most often emitted by combustion sources) and hydrazine (most often released by certain industrial manufacturing processes)—are among the EPA's designated urban air toxics, a listing of 33 priority TACs (plus DPM and coke oven emissions), which are considered to pose significant health risks in urban areas of the United States. Others—which may or may not be important in

the SFBA—include radionuclides (usually from natural sources and controlled medical or testing uses); titanium tetrachloride (from titanium metals manufacturing); and hydrochloric acid, hydrofluoric acid, and sulfuric acid (most often emitted by certain manufacturing processes and refining).

In addition, the chemical speciation profiles that were applied to estimate TAC emissions from pre-existing emission inventories of TOG and PM_{10} contain significant uncertainties. Due to the limitations of available chemical analysis techniques, some of the speciation profiles include large reported proportions of unknown or unidentified species—occasionally as much as 50% to 80% of the total mass—contributing to total TOG or PM emissions. Measurements of source-specific chemical speciation profiles are expensive; therefore, the number of measurements for each source category and TAC are fairly limited in number and may contain inaccuracies or large errors. We are particularly concerned about the large proportion of chromium attributed to the ARB's source profile for construction and demolition activities (as noted in Section 6.1). However, in general, these issues with speciation profiles are difficult and expensive to resolve and will likely take many years of gradual efforts to address.

Finally, there are uncertainties associated with the assumption that 5% of chromium compounds (unspecified), which are associated mostly with fugitive dust emissions, are emitted in their most highly toxic form: chromium VI (or hexavalent chromium). As described in Section 3, speciation profiles containing unspecified chromium were adjusted to reflect this assumption that 5% of the chromium fraction in each profile was made up of chromium VI. Because this assumption was based on ambient measurements taken during the 1980s, we recommend further research into the relative proportions of chromium (III) and chromium (IV) emitted in fugitive dust and other emission sources.

6.1 RECOMMENDATIONS FOR CONTINUING DEVELOPMENT OF THE TOXIC AIR CONTAMINANT EMISSION INVENTORIES

The following strategies are suggested to begin addressing some of the weaknesses in the TAC inventories:

- To the extent feasible, use bottom-up methods for emission inventory development beginning with the highest priority TACs and source types. Prioritize TACs that seem most likely to pose health risks in the SFBA for further emission inventory development. We suggest prioritizing TACs that are listed among the EPA's designated urban air toxics or that appear prominently in Figures 4-1, 4-3, and 4-5. Identify source categories likely to emit prioritized TACs, such as those shown in Figures 4-2, 4-4, and 4-6 or identified in EPA guidance documents (U.S. Environmental Protection Agency, 2005; 1998, Appendix I).
- Use readily available (or, if possible, bottom-up methods) to estimate emissions of TACs that are currently omitted from the inventories: quinoline, hydrazine, radionuclides, titanium tetrachloride, hydrochloric acid, hydrofluoric acid, and sulfuric acid.
- Complete the correction of the coordinates of point sources that were located according to ZIP Code centroids. Of the 137 facilities so located, 24 were included in BAAQMD's

review of spatial information for the most significant point sources (see Section 2.3). A list of remaining point sources located by ZIP Code centroids is provided in Appendix C.

- Use the ARB database of TAC emissions reported for its AB 2588 Air Toxics Hot Spots Program (California Air Resources Board, 2001b) as a means of checking the BAAQMD point source inventory for completeness and accuracy. In a few cases, 1996 emissions reported in the AB 2588 database of point sources in the SFBA greatly exceeded year-2000 emissions included in the BAAQMD's TAC inventory of point sources. These differences may accurately represent reduced levels of TAC emissions; however, it may be worthwhile to verify this information. The AB 2588 database reported 972 tons per year (tpy) of methylene chloride emitted from point sources in the SFBA (compared to 59 tpy reported in the BAAQMD inventory); 5 tpy of manganese (compared to 0.6 tpy); and 110 pounds per year of beryllium (compared to 1 pound per year).
- Investigate the chromium content of ARB's recommended PM_{10} speciation profiles, particularly the profile assigned to construction and demolition dust. Chromium VI is a recognized compound of concern in Portland cement (Klemm, 1994); however, the actual chromium content of construction and demolition dust SFBA counties is unknown. Further research is warranted into the relative proportions of chromium (III) and chromium (IV) emitted in construction and demolition dust and other emission sources.
- Investigate the spatial distribution of emissions from construction activities. Construction dust and construction equipment appear to be significant sources of cancer toxicity-weighted emissions and chronic risk-weighted emissions, but the locations of these emission sources are likely to change with time as construction projects are completed and new projects begin.
- Investigate the assumption that all PM emissions from residual-fueled ships should be classified as DPM, a toxic air contaminant. While this assumption appears reasonable and in keeping with previous ARB studies, further discussions with ship emissions experts may be warranted.

6.2 RECOMMENDATIONS FOR REVIEWING THE CRITERIA POLLUTANT INVENTORIES

The BAAQMD's TOG and PM_{10} emission inventories, which were the basis of the TAC inventories, previously underwent a thorough quality assurance and quality control (QA/QC) review to support of the goals and objectives of the $CCOS$.⁵ In addition, the inventory files received from the BAAQMD were given a cursory QA/QC review prior to the development of the BAAQMD's TAC inventories. A few unusual features were noted and corrected, including a small "hot spot" (about 5 grid cells) of emissions from on-road mobile sources located along a rural stretch of Highway 152 east of Gilroy and north of Hollister. Other features of the inventories might warrant investigation:

• Emissions from on-road mobile sources along Highway 1 in San Mateo County appear to be located approximately 2 km too far west (over Pacific waters).

l ⁵ More information about the CCOS is available at <http://www.arb.ca.gov/airways/ccaqs.htm>.

• We considered whether emissions estimated for ships and commercial boats might be too low for an area with active ports. We verified the reasonableness of the emissions by comparing the BAAQMD's emission inventories to three other inventories—those for Baton Rouge, Louisiana; Houston-Galveston, Texas; and Los Angeles, California. These three port cities were selected for the comparison because we are familiar with the underpinnings of their inventories and are highly confident about the validity of the estimated emissions for commercial marine vessels. In these three cities, commercial marine vessels were estimated to emit 10% to 24% of total $PM_{2.5}$ from all non-road mobile sources: 24% in Baton Rouge, Louisiana; 10% in Houston-Galveston, Texas; and 16% in the Los Angeles area (Reid et al., 2004; California Air Resources Board, 2001a). The BAAQMD's inventory attributes 22% of total non-road PM_{2.5} emissions to commercial marine vessels, which is within the range of values estimated for the other cities.

7. REFERENCES

- California Air Resources Board (1985) Staff report: initial statement of reasons for proposed rule identifying hexavalent chromium as a toxic air contaminant. Prepared by the California Air Resources Board, Sacramento, CA, December.
- California Air Resources Board (2001a) The 2001 California almanac of emissions and air quality. By the Planning and Technical Support Division, Sacramento, California.
- California Air Resources Board (2001b) California toxics inventory. Available on the Internet at <http://www.arb.ca.gov/toxics/cti/cti.htm>.
- Community Action to Fight Asthma (2004) Asthma hospitalization rates (1999-2000, CA OSHPD data) by zip code tabulation area for 15 counties in four regions of California. June.
- Funk T.H., Stiefer P.S., and Chinkin L.R. (2001) Development of gridded spatial allocation factors for the State of California. Technical memorandum prepared for the California Air Resources Board, Sacramento, CA, by Sonoma Technology, Inc., Petaluma, CA, STI-900201/999542-2092-TM, July.
- InfoUSA (2005) Mail list email, business, sales leads and consumer mailing list. Available on the Internet at $\langle \frac{http://www.infousa.com/}{$.
- Klemm W.A. (1994) Hexavalent chromium in Portland cement. Cement, Concrete, and Aggregates 16 (1), 43-47.
- Reid S.B., Sullivan D.C., Penfold B.M., Funk T.H., Tamura T.M., Stiefer P.S., Raffuse S.M., and Arkinson H.L. (2004) Emission inventory development for mobile sources and agricultural dust sources for the Central States. Final report prepared for The Central States Air Resource Agencies and The Central Regional Air Planning Association, Oklahoma City, OK, by Sonoma Technology, Inc., Petaluma, CA, STI-903574-2611-FR, October.
- True North Research (2005) Spare the air tonight study: 2004-2005 winter wood smoke season. Prepared for the Bay Area Air Quality Management District, San Francisco, CA, April.
- U.S. Census Bureau (2000) 1997 economic census of wholesale trade, geographic area series: California. Report by the U.S. Census Bureau, Washington, DC, EC97W42A-CA(RV), March.
- U.S. Environmental Protection Agency (1998) Handbook for air toxics emission inventory development. Office of Air Quality Planning and Standards, EPA-454/B-98-002, November.
- U.S. Environmental Protection Agency (2005) Locating and estimating (L&E) documents. Available on the Internet at <http://www.epa.gov/ttn/chief/le/>.
- U.S. Geological Survey (2001) National Land Cover Dataset (NLCD) products California. Available on the Internet at <http://landcover.usgs.gov/uslandcover.php>.

APPENDIX A

BAAQMD MODELING GRID DEFINITION

BAAQMD Modeling Grid Definition

Lower Left Corner (origin): Y = -385131.6m, X= -302910.3m Grid cell size: 2km x 2km Number of grid cells: 90,000 (300 grid cells x 300 grid cells)

APPENDIX B

SPECIATION OF TOG AND PM EMISSIONS FROM AREA AND NON-ROAD SOURCES AND FROM ON-ROAD MOBILE SOURCES

SPECIATION OF EMISSIONS FROM AREA AND NON-ROAD MOBILE SOURCES

The Bay Area Air Quality Management District (BAAQMD) provided STI with year-2000 county-level emission inventories of TOG and PM_{10} for area and non-road mobile sources. These inventories contain annual average emissions (tons/day) by emission inventory code (EIC). To prepare inventories of toxic air contaminants (TACs), STI assigned speciation profiles to each EIC code. Thus, TOG and PM_{10} emissions were transformed into individual chemical species which allowed specific TACs to be identified. In addition, STI prepared estimates of PM_{2.5} emissions for speciation. (We anticipate that the BAAQMD will model dispersion and deposition of airborne $PM_{2.5}$ differently from that of PM_{10} .)

The California Air Resources Board (ARB) maintains TOG and PM speciation profiles and a cross-reference table, which indicates which TOG or PM profile should be assigned to each EIC code. ARB also provides PM_{10} and $PM_{2.5}$ size fractions for each PM profile. These ARB datasets were the starting points for estimating $PM_{2.5}$ emissions (based on the BAAQMD's PM_{10} inventory and $PM_{2.5}$: PM_{10} ratios) and for speciating the BAAQMD's TOG, PM_{10} , and resultant PM_{2.5} inventories. In most cases, we considered the ARB's recommended speciation profile-to-EIC assignments to be appropriate for use. However, in some cases, ARB did not provide a recommended speciation profile (or composite profile). For these cases, we identified appropriate speciation profiles available from the Desert Research Institute (DRI) or listed in the EPA's Speciate 3.2 database.

Table B-1 lists the EIC codes and speciation profiles for which alternative DRI or EPA profiles were used. In addition, the accompanying Microsoft Excel file contains tables of available speciation profile-to-EIC assignments, complete listings of the TOG and PM profiles recommended for use, and ARB's $PM_{2.5}$ -to- $PM₁₀$ size ratios.

Table B-1. Non-ARB speciation profiles selected for use in speciating the BAAQMD inventories.

B-4

SPECIATION OF TOG AND PM EMISSIONS FROM ON-ROAD MOBILE SOURCES

BAAQMD provided STI with gridded on-road mobile source inventories of TOG and PM₁₀ emissions in Modeling Emissions Data System (MEDS) format. MEDS files were provided for both a July weekday and a January weekday in 2000, and the emissions from these inventories were averaged by grid cell to produce an annualized on-road emissions inventory.

The on-road emission estimates provided by BAAQMD were based on emission factors generated with ARB's EMFAC model, and BAAQMD also provided STI with an input file listing the key settings and input options used to run the EMFAC model. EMFAC, unlike the EPA's MOBILE6 model, does not generate toxic emission factors; therefore, to prepare inventories of toxic air contaminants (TACs), STI assigned speciation profiles to each source category contained in the BAAQMD on-road emissions inventory. Thus, TOG and PM_{10} emissions were transformed into individual chemical species so that specific TACs could be identified. In addition, STI prepared estimates of $PM_{2.5}$ for speciation, anticipating that BAAQMD will model dispersion and deposition of airborne $PM_{2.5}$ differently from that of PM_{10} .

The ARB maintains TOG and PM speciation profiles, as well as PM_{10} and $PM_{2.5}$ size fractions for each PM profile. The ARB has also assembled a cross-reference table that indicates which TOG or PM profile should be assigned to a given Emission Inventory Code (EIC). However, the BAAQMD MEDS inventory lists emissions by a source category code (SCC) numbered from 1 to 12 rather than by EIC. Therefore, STI assigned the most appropriate ARB profile to each SCC code, and these assignments are shown in Table B-2.

In addition, STI ran the EPA's MOBILE6 model using input parameters derived from the EMFAC runs done by BAAQMD. Toxic speciation profiles for the six toxic pollutants covered by MOBILE6 (benzene, 1,3-butadiene, formaldehyde, acetaldehyde, acrolein, and MTBE) were derived from model outputs, and TOG fractions for these TACs were compared to TOG fractions listed in corresponding ARB profiles. Table B-3 shows the results of this comparison, and one can see that the ARB profiles generally predict higher TOG fractions for the six TACs in MOBILE6, with differences being especially pronounced for MTBE (this is likely because of California fuels characteristics being poorly modeled by MOBILE6, a national model).

To adopt a conservative approach, STI decided to use ARB profiles and size fractions for all on-road mobile source speciation. An accompanying Microsoft Excel file contains listings of the TOG and PM profiles recommended for use in speciating the on-road inventory (see Table B-2), as well as ARB's $PM_{2.5}$ -to-PM₁₀ size ratios.

SCC ^a	Description	Profile	Source	Profile Name
1a	Exhaust Particulate Emissions	425	ARB	Diesel Vehicle Exhaust
1 _b	Tire Wear Particulate Emissions	472	ARB	Tire Wear
1c	Brake Wear Particulate Emissions	473	ARB	Brake Wear
$\overline{2}$	Catalyst Start Exhaust	877	ARB	Gasoline - catalyst - FTP Bag 1-3 STARTS - ARB IUS summer 1996
3	Catalyst Running Exhaust	438	ARB	Gasoline - catalyst - stabilized exhaust - ARB IUS summer 1999
$\overline{4}$	Non-catalyst Start Exhaust	402	ARB	Gasoline - non-cat - FTP bag1-3 STARTS - ARB IUS summer 1996
5	Non-catalyst Running Exhaust	401	ARB	Gasoline - non-cat - stabilized exhaust - ARB IUS summer 1996
6	Hot Soak	422	ARB	CBG - hot soak - ARB IUS 1999-2000 - LDV
7	Diurnal Evaporatives	906	ARB	Gasoline - diurnal & resting evaporatives - UCBerk - headspace vapors
8	Diesel Exhaust	818	ARB	Farm equipment - diesel - light and heavy
9	Running Evaporatives	906	ARB	Gasoline - diurnal & resting evaporatives - UCBerk - headspace vapors
10	Resting Evaporatives	906	ARB	Gasoline - diurnal & resting evaporatives - UCBerk - headspace vapors
11	Multi-Day Resting	906	ARB	Gasoline - diurnal & resting evaporatives - UCBerk - headspace vapors
12	Multi-Day Diurnal	906	ARB	Gasoline - diurnal & resting evaporatives - UCBerk - headspace vapors

Table B-2. PM and TOG speciation profile assignments by SCC code.

 a In the BAAQMD MEDS inventory, only TOG emissions are reported in SCC2-SCC12.

B-6

		TOG Percentage									
		Exhaust				Hot Soak		Evaporative			
		ARB	ARB	ARB	ARB	M6	ARB	M ₆	ARB	M6	M ₆
CAS	CHEMICAL NAME	401	402	438	877	Exhaust	422	Hot Soak	906	Diurnal	Resting
100414	ETHYLBENZENE	1.50	1.39	1.09	1.54		0.53		0.11		
100425	STYRENE	0.13	0.14	0.13	0.25		0.02				
106423	P-XYLENE								0.10		
106990	1,3-BUTADIENE	0.83	0.78	0.56	0.70	0.53	0.01				
107028	ACROLEIN (2-PROPENAL)	0.18	0.13	0.14	0.11	0.08					
108383	M & P-XYLENE	8.90	9.50	7.41	10.34		4.02		0.64		
108883	TOLUENE	6.79	7.37	5.99	7.25		3.42		1.59		
110543	N-HEXANE	1.31	1.69	1.61	1.74		1.42		1.44		
123386	PROPIONALDEHYDE	0.13	0.07	0.04	0.06						
1634044	METHYL T-BUTYL ETHER	1.86	3.80	1.97	3.02	0.36	12.66	3.85	16.83	3.24	3.23
	(MTBE)										
50000	FORMALDEHYDE	3.12	1.46	1.73	1.31	1.92					
540841	$2.2.4-$	1.99	1.58	1.75	1.92		1.07		1.21		
	TRIMETHYLPENTANE										
67561	METHANOL	1.40	1.68	0.83	2.46		0.79				
71432	BENZENE	3.44	2.75	2.68	2.47	3.52	0.84	0.43	0.36	0.40	0.40
75070	ACETALDEHYDE	0.75	0.35	0.25	0.40	0.65					
78933	METHYL ETHYL KETONE	0.06	0.10	0.02	0.06						
	(MEK) (2-BUTANONE)										
91203	NAPHTHALENE	0.13	0.02	0.05	0.07		0.00				
95476	O-XYLENE	1.55	1.62	1.29	1.78		0.66		0.12		
98828	(1-METHYLETHYL)	0.10	0.06	0.02	0.12		0.04		0.02		
	BENZENE										

Table B-3. ARB TOG speciation profiles with compared with profiles derived from MOBILE6 outputs.

APPENDIX C

SPATIAL ALLOCATION OF EMISSIONS FROM AREA AND NON-ROAD MOBILE SOURCES

County-level emissions were geographically distributed to the spatial resolution of the BAAQMD modeling domain (2 km x 2 km). Because the exact locations of emissions sources are unknown at this spatial scale, GIS databases with suitable spatial resolutions were used as surrogates to represent the locations of related emissions sources. County-level emissions were allocated to individual grid cells proportionally according to the spatial patterns of the surrogate GIS data.

APPROACH

Allocation factors were developed for individual grid cells by acquiring GIS data and processing them within a customized ArcGIS Visual Basic (VBA) program that outputs allocation factors by grid cell to Microsoft Access database tables. Four basic types of spatial allocation calculations were used to develop the spatial surrogates.

- 1. Total Area—Calculate the fraction of total area that falls within a grid cell (e.g., industrial land use).
- 2. Line Length—Calculate the fraction of total line length that falls within the grid cell.
- 3. Line Density—Calculate the fraction of total line density that falls within the grid cell.
- 4. Point Count—Calculate the fraction of total point locations that falls within the grid cell.

DATA SOURCES

A variety of data sets, listed below, were acquired and used to prepare the surrogate fractions:

- Land Use—The majority of the BAAQMD surrogate development relied on Existing Land Use–2000 data set developed by the Association of Bay Area Governments (ABAG). The Existing Land Use–2000 data set incorporates the U.S. Geological Survey National Land Cover Dataset (NLCD) and county assessors' information. Land use categories are broken down into four levels of numerical subdivisions, which are used to denote the levels of detail in the map data. The categories used by ABAG are described in Existing Land Use–2000 documentation. The individual surrogates were created by determining the fractions of total area of selected land use categories that fell within individual grid cells.
- Land Cover—A satellite-derived land cover database for northern California was acquired from the NLCD in a Geo-TIFF format. Each pixel of the 30-m resolution NLCD was assigned one of the 38 land cover classes. These data were created in the early 1990s and are the most complete land cover data set readily available (U.S. Geological Survey, 2001).
- InfoUSA Business Locations—For a handful of activities (e.g., marine shops, car dealerships, carwash centers, etc.), surrogates were developed using business addresses purchased from InfoUSA (InfoUSA, 2005). Addresses were acquired for businesses within the BAAQMD domain. The addresses were geocoded and assigned to grid cells in the domain. The spatial surrogate data were created by determining the fractions of total businesses that fell within individual grid cells.
- U.S. Electronic Yellow Pages (*ProCD Select Phone*) Business Locations—For a handful of activities, surrogates were used from the Central California Ozone Study (CCOS) gridded surrogates project (Funk et al., 2001). Business addresses for such categories as auto body shops, dry cleaners, restaurants, gas stations, and wineries were acquired from the U.S. Electronic Yellow Pages. The addresses were then geocoded and assigned to grid cells in the domain. The surrogates were created by determining the fractions of

total businesses that fell within individual grid cells or combined with other surrogate types.

- Railroads—Year-2002 railroad tracks were acquired from the U.S. Bureau of Transportation Statistics in Shapefile format. Tracks are represented as linear segments. The data include terminal and rail yard lines in addition to mainline tracks. The proportions of total track length within individual grid cells were used to develop surrogates for rail-related activities.
- Oil and Gas Wells—Onshore and offshore oil and gas well locations were gathered as part of the CCOS gridded surrogates project (Funk et al., 2001). For all areas of California, including the Bay Area, abandoned wells were filtered out and all other well sites—active, completed, idle, directional, etc.—were retained. The oil and gas wells within the BAAQMD domain were queried out of the CCOS gridded surrogate database and assigned to BAAQMD grid cells.
- Landfills—Geographic coordinates of landfill locations were gathered from the Integrated Waste Management Board as part of the CCOS gridded surrogates project (Funk et al., 2001). Only sites flagged as active by the data provider were included. Landfills within the BAAQMD domain were queried out of the CCOS gridded surrogate database and assigned to BAAQMD grid cells.
- Sand and Gravel Mines—Geographic coordinates of sand and gravel mines were gathered as part of the CCOS gridded surrogates project from the National Atlas database (Funk et al., 2001). The sand and gravel mines within the BAAQMD domain were queried out of the CCOS gridded surrogate database and assigned to BAAQMD grid cells.
- Road Activity—The TeleAtlas MultiNet road network was used to create line-density surrogates for some highway and traffic-related emission sources. Because traffic count information was not provided, line density surrogates were developed based on the functional class definition (FRC code). Thus, a major freeway (FRC code = 0) has a higher weighting scheme than a local road (FRC code $= 6$).
- Recreational Water—Water body polygons were derived from the 2000 Census data for the entire state as part of the CCOS gridded surrogates project (Funk et al., 2001). The water body file was edited in order to eliminate dry lakes and to retain lakes and rivers with boat marinas based on information published by the California Water Resources Agency, Department of Boating and Waterways. The water bodies within the BAAQMD domain were queried out of the CCOS gridded surrogate database and assigned to BAAQMD grid cells.
- Shipping Lanes—Year-2000 shipping lanes were acquired from the U.S. Bureau of Transportation Statistics in Shapefile format. Shipping lanes are represented as linear segments. The data include shipping lanes both within the San Francisco Bay and the coastal region beyond the nine Bay Area counties. The proportions of total shipping lane length within individual grid cells were used to develop surrogates for shipping-related activities.

SPECIAL CASES

Computed surrogates were developed for construction activities. Because construction activities are transient (they do not take place in a fixed location), developing spatial allocation factors for these sources was more challenging. Cities tend to expand outward, and new building construction usually occurs in the suburban and rural regions surrounding the central business district, though some fill-in development does occur, as well as maintenance and remodeling on existing structures. As part of the CCOS gridded surrogates project (Funk et al., 2001), an approach to compute spatial surrogates for construction emission was developed. The resulting computed surrogates were applied to the BAAQMD domain.

Spatial allocation factors for building construction activities (both residential and nonresidential) were computed by taking the difference in each grid cell for retail, non-retail, and housing data between a past year and a future year (e.g., 1995 and 2000, respectively) to obtain the differences between past-year and future-year number of units. This calculation was used to determine where new construction activity was likely to occur. In addition to determining where new construction occurred, repair and maintenance of existing structures was also considered.

According to the U.S. Census Bureau's 1997 Economic Census Construction Report (U.S. Census Bureau, 2000), approximately 30% of total residential construction expenditures in the State of California are attributed to repair and maintenance of existing residences. Because new construction activity tends to be more equipment-intensive and, consequently, produces more emissions, areas where new construction occurred (based on the difference calculation) were weighted more heavily than areas where maintenance and repair occurred. Equation C-1 shows an example calculation.

Computed Surrogate for Residential Construction between years A and B:

$$
S_{rc(x)} = (TH_{A(x)} \times 0.30) + (TH_{B(x)} - TH_{A(x)})
$$
 (C-1)

where:

 $S_{\text{rc}(x)}$ = residential construction surrogate for grid cell "x" $TH_{A(x)}$ = total housing value in year A for grid cell "x" $TH_{B(x)}$ = total housing value in year B for grid cell "x" 0.30 = percentage of total statewide residential construction expenditure spent on repair and maintenance

Table C-1 depicts the spatial surrogates used to allocate area and non-road mobile sources for the BAAQMD domain. In order to assign a geographical location and a grid cell identifier to individual point sources, facility addresses reported in the BAAQMD's point source inventory were geocoded. Each point source in the BAAQMD's inventory file was assigned a geographic location and a geocode match type field. A match type of "1" is the best match type possible, where the geographic location is accurate to an exact house number within a single side of a single street block. A match type of "4" is accurate only to the 5-digit zip code centroid. Of 3,359 unique addresses in the BAAQMD's point source inventory file, 4% received a match type "4" and 96% received a match type "1". Table C-2 depicts the unique point source addresses that received a match type of "4" along with there corresponding TAC emissions estimates.

Table C-1. Summary of spatial surrogates used to allocate area and non-road mobile sources for the BAAQMD domain.

^a See memorandum for proposed spatial approach for documentation of surrogate numerical codes.

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^a See memo for proposed spatial approach for documentation of surrogate numerical codes.

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Plant Number ^a	Plant Name	SIC	Address	City	Total HAP Emissions $\left(1bs/yr\right)^b$	Longitude	Latitude	# of individual point sources ^c
1534	South Bayside System Authority	4952	Radio Road, End of	Redwood City	39,111	-122.2459	37.53565	13
4116	San Francisco, City & County, PUC	4952	3500 Great Highway	San Francisco	8,521	-122.49742	37.71745	9
12089	Precision Cabinets & Trim	2434	700 Hrvst Prk Drv St	Brentwood	3,547	-121.70196	37.93333	2
11671	Gas Recovery Systems, Inc	4911	Landfill, American Canyon	Napa	3,324	-122.29874	38.32516	26
1204	Cultured Stone Corporation	3299	Highway 29 & Tower Rd	Napa	3,278	-122.28519	38.29346	3
3232	OEA Areospace Inc	3728	Explosive Tech Rd, at Highway 12	Fairfield	2,848	-122.03037	38.26551	11
653	Central Marin Sanitation Agency	4952	Anderson Drive, East end	San Rafael	2,520	-122.51632	37.9714	15
3029	Contra Costa Sanitary Landfill	4953	James Donlon Blyd	Antioch	2,067	-121.81141	37.99963	29
3256	Turk Island Solid Waste Disposal Site	4953	Union City Boulevard	Union City	1,654	-122.04476	37.59136	16
4094	San Quentin State Prison	9223	CA State Prison	San Quentin	1,432	-122.48303	37.94277	17
5264	Mc Clellan Square Cleaners	7216	10477 De Anza Blvd	Cupertino	1,349	-122.03786	37.32082	$\overline{1}$

Table C-2. Total HAP emissions from point source (plant) addresses that geocoded to a zip code centroid match (match-type 4).

^a Plant number based on point of source inventory file provided by BAAQMD.

^b TAC emissions summed by plant address.

^c Total number of individual point sources located at the associated plant address.

Table C-2. Total HAP emissions from point source (plant) addresses that geocoded to a zip code centroid match (match-type 4).

 α Plant number based on point of source inventory file provided by BAAQMD.

TAC emissions summed by plant address.

 \overline{c} Total number of individual point sources located at the associated plant address.

C-10

 α Plant number based on point of source inventory file provided by BAAQMD.

TAC emissions summed by plant address.

c
Total number of individual point sources located at the associated plant address.

C-11

 α Plant number based on point of source inventory file provided by BAAQMD.

TAC emissions summed by plant address.

 \degree Total number of individual point sources located at the associated plant address.

^a Plant number based on point of source inventory file provided by BAAQMD.

^b TAC emissions summed by plant address.

^c Total number of individual point sources located at the associated plant address.

 α Plant number based on point of source inventory file provided by BAAQMD.

TAC emissions summed by plant address.

 \degree Total number of individual point sources located at the associated plant address.

 α Plant number based on point of source inventory file provided by BAAQMD.

TAC emissions summed by plant address.

 \overrightarrow{c} Total number of individual point sources located at the associated plant address.

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Plant Number ^a	Plant Name	SIC	Address	City	Total HAP Emissions $\left(lbs/yr\right)^b$	Longitude	Latitude	# of individual point sources ^c
7852	California Water Service Company	9999	Moore Rd, Station 5 F	Atherton		-122.18216	37.45103	8
10211	Mega Sand Inc	4953	Decker Island	Antioch		-121.81141	37.99963	10
10834	Pacific Custom Materials Inc	3299	9000 Carquinez Scenic Dr	Port Costa		-122.18824	38.045	$\overline{2}$
128	Syar Industries, Inc.	2951	Lake Herman Road	Vallejo	1	-122.21114	38.10168	5
3194	City of Alameda, Maint Serv Center	4953	Doolittle Drive	Alameda		-122.2589	37.76884	26
12181	American Airlines	4522	Spr By Hngr Cst Rd, Bldg #1060	San Francisco	1	-122.39608	37.63612	10
9327	Napa Pipe Corporation	3479	Hwy 29 & Kaiser Rd	Napa		-122.29874	38.32516	5
11827	Delta c/o Environmental Cost Management	9999	SF Int'l Airport, Boarding Area	San Francisco		-122.39608	37.63612	2
896	Hanson Aggregates	2951	Pine Hollow Road	Clayton		-121.9213	37.93193	$\overline{4}$
479	Treasure Island	9711	Treasure Island	San Francisco	1	-122.3727	37.824	\mathfrak{Z}
			Total HAP emissions by match-type "4" locations	$811,371^d$	Total individual point sources at match-type "4" locations	3,977		

Table C-2. Total HAP emissions from point source (plant) addresses that geocoded to a zip code centroid match (match-type 4).

 α Plant number based on point of source inventory file provided by BAAQMD.

TAC emissions summed by plant address.

 \degree Total number of individual point sources located at the associated plant address.

^d This figure corresponds to 7.5% of total mass-based TAC emissions, and 11.9% of TAC emissions from point sources.

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REFERENCES

- Funk T.H., Stiefer P.S., and Chinkin L.R. (2001) Development of gridded spatial allocation factors for the State of California. Technical memorandum prepared for the California Air Resources Board, Sacramento, CA, by Sonoma Technology, Inc., Petaluma, CA, STI-900201/999542-2092-TM, July.
- InfoUSA (2005) Mail list email, business, sales leads and consumer mailing list. Available on the Internet at <http://www.infousa.com/>.
- U.S. Census Bureau (2000) 1997 economic census of wholesale trade, geographic area series: California. Report by the U.S. Census Bureau, Washington, DC, EC97W42A-CA(RV), March.
- U.S. Geological Survey (2001) National Land Cover Dataset (NLCD) products California. Available on the Internet at <http://landcover.usgs.gov/uslandcover.php>.

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APPENDIX D

SELECTION AND APPLICATION OF CANCER UNIT RISK FACTORS AND NON-CANCER REFERENCE CONCENTRATIONS

Sonoma Technology, Inc. (STI) applied TAC-specific cancer unit risk factors (UR) and non-cancer Reference Concentrations (RfC) to estimated TAP emissions. UR factors and RfCs were acquired from ARB in conjunction with California EPA's Office of Environmental Health Hazard Assessment (OEHHA), EPA's Integrated Risk Information System (IRIS), and EPA's Technology Transfer Network, along with secondary sources used to estimate factors for important TAPs not reported by these sources. A table of primary and alternative factors was compiled using the following hierarchy of preference (see Table D-1 for full citations):

- 1. Consolidated Table of OEHHA/ARB Approved Risk Assessment Health Values.
- 2. EPA OEHHA September 2004—OEHHA Cancer Potency List.
- 3. EPA Integrated Risk Information System (IRIS).
- 4. EPA Technology Transfer Network, Table 1 Prioritized Chronic Dose-Response Values for Screening Risk Assessments (2/28/05) (TTN).

The following sources were used when primary or alternative factors were not available from OEHHA/ARB, IRIS, or TTN:

- 1. EPA Environmental Information Management System, Trichloroethylene Health Risk Assessment: Synthesis and Characterization (External Review Draft), v1.0, EPA/600/P-01/002A, 2001.
- 2. Risk Assessment Information System (RAIS), Chemical-Specific Toxicity Values.

In addition, the following tables were used to cross-reference dioxin and furan descriptors for which Chemical Abstracts Service numbers (CAS) have not yet been assigned:

- 1. California Air Resources Board (ARB), California Air Toxics Emission Factors, Database User's Manual Version 1.2, Table 9 Listing of Field Values for the Field Substance.
- 2. South Coast Air Quality Management (SCAQMD), Risk Assessment Procedures for Rules 1401 and 212, Attachment C, Table – 8 Unit Risk Factor (U), Reference Exposure Level (REL) and Multi Pathway Adjustment Factors (MP).

Selection codes 1, 2, and 3 were used to designate Selected (1) and Alternative (2 or 3) factors for each of the risk categories evaluated: Inhalation Cancer, Chronic Inhalation, and Acute Inhalation. Selection codes were assigned according to the hierarchy of preference described above. Finalized factors were used to risk-weight emission estimates. When factors were given as a range (e.g., Benzene, CAS# 71432, IRIS), selection codes were assigned to each end point of the range, with the higher selection preference given to the high end of the range. By applying the high end of the range, the risk-weighted inventories represent a worst-case, conservatively high-risk scenario. A condensed table of UR factors and RfCs is provided in Table D-2. Table D-3 lists range-related and other uncertainty issues for affected TACs and tabulates associated risk-weighted emissions for each issue.

Please note the following information about substituted CAS number assignments of the TAC inventories:

- A CAS number of 111159 was used to designate emissions of "glycol ethers (other/not specified)" which were reported in the BAAQMD's inventory of point sources. CAS 111159 normally denotes cellosolve acetate, which is a type of glycol either.
- The CAS number 1746016 normally designates 2,3,7,8-tetrachlorodibenzo-p-dioxin. It was applied as such for the TAC inventories for area, non-road mobile, and/or on-road mobile sources. In addition, this CAS number (1746016) was used to designate emissions of "undifferentiated chlorinated dioxins and furans" reported in the BAAQMD's inventory of point sources.

a Acute reference concentrations are generally based on a one-hour exposure.
^b Chronic reference concentrations are based on exposures of a greater duration than 12% of a lifetime of 70 years, so human exposures of more considered chronic.

 \textdegree Cancer risks are calculated based on a lifetime exposure duration (70 years)

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100414 Ethyl Benzene 100414 and a streamer and a streamer ally based on a one-hour exposure.

⁸ Chronic reference c are considered chronic.

 \textdegree Cancer risks are calculated based on a lifetime exposure duration (70 years)

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^a Acute reference concentrations are generally based on a one-hour exposure.
^b Chronic reference concentrations are based on exposures of a greater duration than 12% of a lifetime of 70 years, so human exposures of mor considered chronic.

 \textdegree Cancer risks are calculated based on a lifetime exposure duration (70 years)

Table D-2. Unit risk factors and reference concentrations ("tblUnitRisk").

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^a Acute reference concentrations are generally based on a one-hour exposure.
^b Chronic reference concentrations are based on exposures of a greater duration than 12% of a lifetime of 70 years, so human exposures of mor considered chronic.

 \textdegree Cancer risks are calculated based on a lifetime exposure duration (70 years)

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Table D-2. Unit risk factors and reference concentrations ("tblUnitRisk").

CAS Chemical Acute Inhalation $(\mu g/m^3)$ SelectionCode Acute Inhalation Chronic Inhalation $(\mu g/m^3)$ Selection Code Chronic Inhalation Inhalation Cancer UnitRisk $(\mu g/m^3)^{-1}$ Selection Code Inhalation CancerUR CitationID CitationDate Source 109864 Ethylene Glycol Methyl (2.0E+01 2) 2.0E+01 2 3/27/05 EPA IRIS 110496 Ethylene Glycol Monomethyl 9.0E+01 1 1 1 1 OEHHA/ARB 110805 Cellosolve Solvent 3.7E+02 1 7.0E+01 1 1 5/27/05 OEHHA/ARB 110805 Cellosolve Solvent 2.0E+02 2 2 3/27/05 EPA IRIS 111159 Cellosolve Acetate 1.4E+02 1 3.0E+02 1 1 5/27/05 OEHHA/ARB 74839 | Methyl Bromide | 3.9E+03 | 1 | 5.0E+00 | 1 | | | | | | | | | | | | | 5/27/05 | OEHHA/ARB 74839 Methyl Bromide 5.0E+00 2 5/27/05 EPA IRIS 68122 N,N-Dimethylformamide | | 8.0E+01 | 1 | | | 5/27/05 OEHHA/ARB 68122 N,N-Dimethylformamide 3.0E+01 2 2 5/27/05 EPA IRIS 123911 | p-Dioxane | 3.0E+03 | 1 | 3.0E+03 | 1 | 7.7E-06 | 1 | 1 | 5/27/05 | OEHHA/ARB 123911 p-Dioxane 7.7E-06 5 5/27/05 TTN 121448 Triethylamine | 2.8E+03 | 1 | 2.0E+02 | 1 | | | | | | | | | | 5/27/05 | OEHHA/ARB 121448 Triethylamine 7.0E+00 2 2 5/27/05 EPA IRIS OEHHA/ARB 75354 Vinylidene Chloride 1 7.0E+01 1 1 1 5/27/05 75354 Vinylidene Chloride 2.0E+02 2 2 5/27/05 EPA IRIS 75354 Vinylidene Chloride 5.0E-05 1 6 6/1/05 RAIS 542756 1,3-Dichloropropene 1.6E-05 1 3 5/27/05 OEHHA 542756 1,3-Dichloropropene 2.0E+01 1 4.0E-06 2 2 6/1/05 EPA IRIS 112345 Diethylene Glycol (Monobutyl Ether 2.0E+01 1 1 5 5/27/05 TTN 7440611 Uranium 3.0E-01 1 5 5/27/05 TTN 1582098 Trifluralin 2.2E-06 1 5 5/27/05 TTN

^a Acute reference concentrations are generally based on a one-hour exposure.

 b^b Chronic reference concentrations are based on exposures of a greater duration than 12% of a lifetime of 70 years, so human exposures of more than 8 years are considered chronic.

 \textdegree Cancer risks are calculated based on a lifetime exposure duration (70 years)

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^a Acute reference concentrations are generally based on a one-hour exposure.
^b Chronic reference concentrations are based on exposures of a greater duration than 12% of a lifetime of 70 years, so human exposures of mor considered chronic.

 \textdegree Cancer risks are calculated based on a lifetime exposure duration (70 years)

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^b Chronic reference concentrations are based on exposures of a greater duration than 12% of a lifetime of 70 years, so human exposures of mor considered chronic.

 \textdegree Cancer risks are calculated based on a lifetime exposure duration (70 years)

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Table D-2. Unit risk factors and reference concentrations ("tblUnitRisk").

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 \textdegree Cancer risks are calculated based on a lifetime exposure duration (70 years).

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Table D-3. List of range-related and other risk factor uncertainties.

aSelection codes of "2" or "3" were assigned to the listed UR factors or RfCs; therefore, the reported uncertainties do not directly affect the TAC inventories prepared for the BAAQMD. Only if selection codes are revised to apply the listed UR factors or RfCs (i.e., selection codes of "1" assigned) will the reported uncertainties directly affect the TAC inventories.

 b "The unit risk should not be used if the air concentration exceeds 4 μ g/m³, since above this concentration the unit risk may not be appropriate." (IRIS).

^d "These subchronic and chronic non-cancer toxicity values are found in Agency documents, but were calculated by alternative methods that are not currently practiced by the RfD/RfC Work Group. These values are considered to be adequate provisional values for risk assessment purposes at Superfund and RCRA sites, but are subject to be reviewed by the RfD/RfC Work Group and revised when necessary to reflect current work group practices." (RAIS).

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^c "The Ethylbenzene cancer toxicity values have been withdrawn by NCEA. This chemical is now being reassessed for IRIS which automatically flags further use of any provisional cancer or non-cancer assessments. However, the RAIS has retained these values and added appropriate footnotes." (RAIS).

^a Selection codes of "2" or "3" were assigned to the listed UR factors or RfCs; therefore, the reported uncertainties do not directly affect the TAC inventories prepared for the BAAQMD. Only if selection codes are revised to apply the listed UR factors or RfCs (i.e., selection codes of "1" assigned) will the reported uncertainties directly affect the TAC inventories.

 b "The unit risk should not be used if the air concentration exceeds 4 μ g/m³, since above this concentration the unit risk may not be appropriate." (IRIS).

^c "The Ethylbenzene cancer toxicity values have been withdrawn by NCEA. This chemical is now being reassessed for IRIS which automatically flags further use of any provisional cancer or non-cancer assessments. However, the RAIS has retained these values and added appropriate footnotes." (RAIS).

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TABLE STRUCTURES, FILE DOCUMENTATION, AND PROCEDURAL STEPS

Table D-4 documents the tables structure for the compilation of cancer URs and noncancer RfCs by citation source to form "tblUnitRisk".

Table D-4. Table structure for "tblUnitRisk".

Selection code queries were developed individually from the above table, then rejoined to form a complete table, "CancerAndNonCancerUnitRiskFactors," for use in estimating riskweighted emissions. Additional fields combining URs or RfCs with selection codes 1, 2, or 3 were included for use in estimating risk-weighted emissions where comparisons between primary and alternative factors are desired. For example, field "CancerUR1" contains UR factors with a selection code of 1 only, field "CancerUR1,2" contains UR factors with selection codes 1 or 2 where selection code 2 UR factors replace selection code 1 UR factors where available, etc. Selection code 1 denotes the primary factor, 2 and 3 denote alternative factors. Additional alternative factors can be added by beginning with selection code 4, etc. (additional selection code queries will need to be added along with revising summary queries).

Table D-5 documents the table structure for the complete table of primary and alternative factors for use in risk-weighting emission.

Table D-5. Table structure for "CancerAndNonCancerUnitRiskFactors".

Table D-6 provides the complete table of primary and alternative UR factors and RfCs for use in toxicity-weighting emissions.

Toxicity-weighted emissions for cancer due to inhalation exposure are calculated according to Equation D-1.

$$
Emissions \times UR_i \div UR_X = Toxicity-Weighted Emissions
$$
 (D-1)

where:

Toxicity-Weighted Emissions = Equivalent emissions of hypothetical compound "X", which would be expected to pose a risk equal to that of the mass of the TAC species i ; equivalent lbs/day

Toxicity-weighted emissions for acute or chronic effects due to inhalation exposure are calculated according to Equation D-2.

$$
Emissions \div RfC_i \times RfC_Y = Toxicity-Weighted Emissions
$$
 (D-2)

where:

Emissions $=$ Mass-based emissions of a TAC species i; lbs/day RfCi = RfC for TAC species i; $\mu g/m^3$ RfCY $=$ RfC for hypothetical compound "Y" $\equiv 1 \mu g/m^3$

Toxicity-Weighted Emissions = Equivalent emissions of hypothetical compound "Y", which would be expected to pose a risk equal to that of the mass of the TAC species i; equivalent lbs/day

Table D-6. Primary and alternative factors for use in toxicity-weighted emission comparisons ("CancerAndNonCancerUnitRiskFactors").

Table D-6. Primary and alternative factors for use in toxicity-weighted emission comparisons ("CancerAndNonCancerUnitRiskFactors").

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CAS	Chemical	CancerUR	CancerUR 1,2	CancerUR 1,3	ChronicRfC	ChronicRfC 1,2	ChronicRfC 1,3	AcuteRf	AcuteRfC 1,2	AcuteRfC 1,3
40321764	$1,2,3,7,8$ - Pentachlorodibenzo-p- Dioxin	3.8E+01	3.8E+01	$3.8E + 01$	4.0E-05	4.0E-05	4.0E-05			
19408743	$1,2,3,7,8,9-$ Hexachlorodibenzo-p- Dioxin	$3.8E + 00$	$1.3E + 00$	$3.8E + 00$	4.0E-04	4.0E-04	4.0E-04			
39227286	$1,2,3,4,7,8$ - Hexachlorodibenzo-p- Dioxin	$3.8E + 00$	$3.8E + 00$	$3.8E + 00$	4.0E-04	4.0E-04	4.0E-04			
55673897	1,2,3,4,7,8,9- Heptachlorodibenzofuran	3.8E-01	3.8E-01	3.8E-01	4.0E-03	4.0E-03	4.0E-03			
51207319	$2,3,7,8-$ Tetrachlorodibenzofuran	$3.8E + 00$	$3.8E + 00$	$3.8E + 00$	4.0E-04	4.0E-04	4.0E-04			
70648269	$1,2,3,4,7,8$ - Hexachlorodibenzofuran	$3.8E + 00$	$3.8E + 00$	$3.8E + 00$	4.0E-04	4.0E-04	4.0E-04			
72918219	$1,2,3,7,8,9-$ Hexachlorodibenzofuran	$3.8E + 00$	$3.8E + 00$	$3.8E + 00$	4.0E-04	4.0E-04	4.0E-04			
67562394	$1,2,3,4,6,7,8$ - Heptachlorodibenzofuran	3.8E-01	3.8E-01	3.8E-01	4.0E-03	4.0E-03	4.0E-03			
106898	1 -Chloro-2,3- Epoxypropane	$2.3E-05$	1.2E-06	$2.3E-05$	$3.0E + 00$	$1.0E + 00$	$3.0E + 00$	$1.3E + 03$	$1.3E + 03$	$1.3E + 03$
140885	Ethyl Acrylate	1.4E-05	1.4E-05	1.4E-05	$4.8E + 01$	$4.8E + 01$	$4.8E + 01$			
75003	Ethyl Chloride				$3.0E + 04$	$1.0E + 04$	$3.0E + 04$			
100414	Ethyl Benzene	1.1E-06	1.1E-06	1.1E-06	$2.0E + 03$	$1.0E + 03$	$2.0E + 03$			
106934	Ethylene Dibromide	7.1E-05	6.0E-04	3.0E-04	8.0E-01	$9.0E + 00$	8.0E-01			
107062	Ethylene Dichloride	2.1E-05	2.6E-05	2.1E-05	$4.0E + 02$	$2.4E + 03$	$4.0E + 02$			
107211	Ethylene Glycol				$4.0E + 02$	$4.0E + 02$	$4.0E + 02$			
75218	Ethylene Oxide	8.8E-05	8.8E-05	8.8E-05	$3.0E + 01$	$3.0E + 01$	$3.0E + 01$			
75343	Ethylidene Dichloride $(1,1-Dichloroethane)$	1.6E-06	1.6E-06	1.6E-06	5.0E+02	$5.0E + 02$	$5.0E + 02$			

Table D-6. Primary and alternative factors for use in toxicity-weighted emission comparisons ("CancerAndNonCancerUnitRiskFactors").

										Page 4 of 6
CAS	Chemical	CancerUR	CancerUR	CancerUR	ChronicRfC	ChronicRfC	ChronicRfC	AcuteRf	AcuteRfC	AcuteRfC
			1,2	1,3		1,2	1,3		1,2	1,3
50000	Formaldehyde	6.0E-06	1.3E-05	5.5E-09	$3.0E + 00$	$9.8E + 00$	$3.0E + 00$	$9.4E + 01$	$9.4E + 01$	$9.4E + 01$
111159	Cellosolve Acetate				$3.0E + 02$	$3.0E + 02$	$3.0E + 02$	$1.4E + 02$	$1.4E + 02$	$1.4E + 02$
110496	Ethylene Glycol Monomethyl Ether Acetate				$9.0E + 01$	$9.0E + 01$	$9.0E + 01$			
112345	Diethylene Glycol Monobutyl Ether				$2.0E + 01$	$2.0E + 01$	$2.0E + 01$			
110805	Cellosolve Solvent				7.0E+01	$2.0E + 02$	7.0E+01	$3.7E + 02$	$3.7E + 02$	$3.7E + 02$
107982	Propylene Glycol Monoethyl Ether				$7.0E + 03$	$7.0E + 03$	$7.0E + 03$			
111900	Diethylene Glycol Monoethyl Ether				$3.0E + 00$	$3.0E + 00$	$3.0E + 00$			
111762	Butyl Cellosolve				$2.0E + 01$	$1.3E + 04$	$2.0E + 01$	$1.4E + 04$	$1.4E + 04$	$1.4E + 04$
109864	Ethylene Glycol Methyl Ether				$6.0E + 01$	$2.0E + 01$	$6.0E + 01$	$9.3E + 01$	$9.3E + 01$	$9.3E + 01$
118741	Hexachlorobenzene	5.1E-04	4.6E-04	5.1E-04	$2.8E + 00$	$3.0E + 00$	$2.8E + 00$			
822060	Hexamethylene Diisocyanate				1.0E-02	1.0E-02	1.0E-02			
110543	Hexane				$7.0E + 03$	$2.0E + 02$	$7.0E + 03$			
7647010	Hydrochloric Acid				$9.0E + 00$	$2.0E + 01$	$9.0E + 00$	$2.1E + 03$	$2.1E + 03$	$2.1E + 03$
7664393	Hydrogen Fluoride				$1.4E + 01$	$1.4E + 01$	$1.4E + 01$	$2.4E + 02$	$2.4E + 02$	$2.4E + 02$
78591	Isophorone	2.7E-07	2.7E-07	2.7E-07	$2.0E + 03$	$2.0E + 03$	$2.0E + 03$			
7439921	Lead	1.2E-05	1.2E-05	1.2E-05	$1.5E + 00$	$1.5E + 00$	$1.5E + 00$			
108316	Maleic Anhydride				7.0E-01	7.0E-01	7.0E-01			
7439965	Manganese				2.0E-01	5.0E-02	2.0E-01			
7439976	Mercury				9.0E-02	3.0E-01	9.0E-02	$1.8E + 00$	$1.8E + 00$	$1.8E + 00$
67561	Methanol				$4.0E + 03$	$4.0E + 03$	$4.0E + 03$	$2.8E + 04$	$2.8E + 04$	$2.8E + 04$
74839	Methyl Bromide				$5.0E + 00$	$5.0E + 00$	$5.0E + 00$	3.9E+03	$3.9E + 03$	$3.9E + 03$
74873	Methyl Chloride	1.8E-06	1.8E-06	1.8E-06	$9.0E + 01$	$9.0E + 01$	$9.0E + 01$			
71556	Methyl Chloroform				$1.0E + 03$	$1.0E + 03$	$1.0E + 03$	$6.8E + 04$	$6.8E + 04$	$6.8E + 04$
78933	Methyl Ethyl Ketone				$1.0E + 03$	$5.0E + 03$	$1.0E + 03$	$1.3E + 04$	$1.3E + 04$	$1.3E + 04$

Table D-6. Primary and alternative factors for use in toxicity-weighted emission comparisons ("CancerAndNonCancerUnitRiskFactors").

Table D-6. Primary and alternative factors for use in toxicity-weighted emission comparisons ("CancerAndNonCancerUnitRiskFactors").

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CAS	Chemical	CancerUR	CancerUR	CancerUR	ChronicRfC	ChronicRfC	ChronicRfC	AcuteRf	AcuteRfC	AcuteRfC
			1,2	1,3		1,2	1,3		1,2	1,3
121448	Triethylamine				$2.0E + 02$	$7.0E + 00$	$2.0E + 02$	$2.8E + 03$	$2.8E + 03$	$2.8E + 03$
1582098	Trifluralin	$2.2E-06$	$2.2E-06$	$2.2E-06$						
108054	Vinyl Acetate				$2.0E + 02$	$2.0E + 02$	$2.0E + 02$			
75014	Vinyl Chloride	7.8E-05	8.8E-06	4.4E-06	$2.6E + 01$	$1.0E + 02$	$2.6E + 01$	$1.8E + 05$	1.8E+05	1.8E+05
75354	Vinylidene Chloride	5.0E-05	5.0E-05	5.0E-05	$7.0E + 01$	$2.0E + 02$	$7.0E + 01$			
106423	p -Xylene				$7.0E + 02$	$1.0E + 02$	$7.0E + 02$	$2.2E + 04$	$2.2E + 04$	$2.2E + 04$
108383	m-Xylene				$7.0E + 02$	$1.0E + 02$	$7.0E + 02$	$2.2E + 04$	$2.2E + 04$	$2.2E + 04$
1330207	Xylenes (Mixture of o,				$7.0E + 02$	$1.0E + 02$	$7.0E + 02$	$2.2E + 04$	$2.2E + 04$	$2.2E + 04$
	m, and p Isomers)									
95476	o-Xylene				$7.0E + 02$	$1.0E + 02$	$7.0E + 02$	$2.2E + 04$	$2.2E + 04$	$2.2E+04$

Table D-6. Primary and alternative factors for use in toxicity-weighted emission comparisons ("CancerAndNonCancerUnitRiskFactors").

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